

On the way to an energy producing nutrient removal WWT plant

Presentation at

Tokyo University

July 16, 2015

Prof. Helmut Kroiss, IWA President
Vienna University of Technology



Where I come from



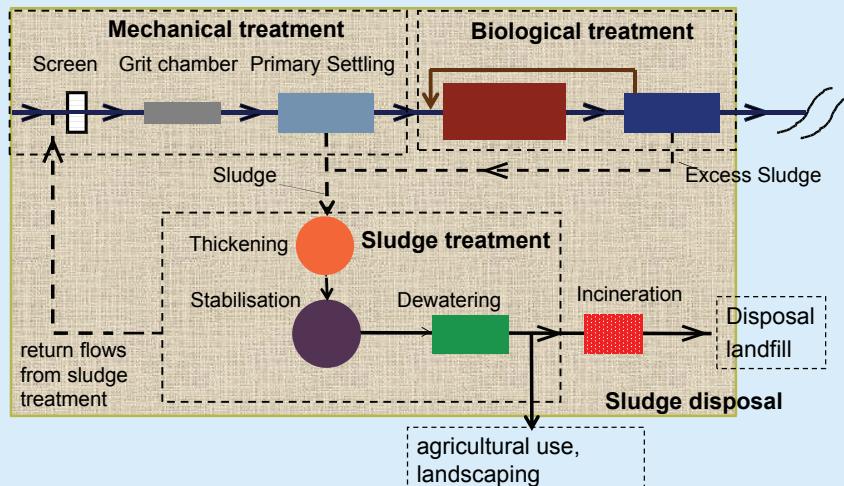
Relevant “primary” power data in kW per inhabitant

- Global Solar power ~10,000
- Global primary power input 3
- EU primary power input 6
- **Municipal Waste water treatment <0.02**
- Waste water pollution ~0.04



Important basics for this presentation

- “Efficient waste water treatment” (AS)
 - ⇒ Full nitrification at any time ($\text{NH}_4\text{-N} < 5 \text{ mg/l}$)
 - ⇒ COD removal >90 % ($\text{BOD}_5 \text{ effluent} < 10 \text{ mg/l}$)
 - ⇒ Phosphorus removal > 85%
 - ⇒ Nitrogen removal >75%
- Why:
 - ⇒ No oxygen removal from receiving water
 - ⇒ Most of micro-pollutants are removed
 - ⇒ Eutrophication abatement (P and N)
 - ⇒ P resource recovery
 - ⇒ Minimisation of energy requirement (N)



Austrian process benchmarking results

Stefan Lindtner Consultant
Austrian Water and Waste Association
<http://www.abwasserbenchmarking.at>

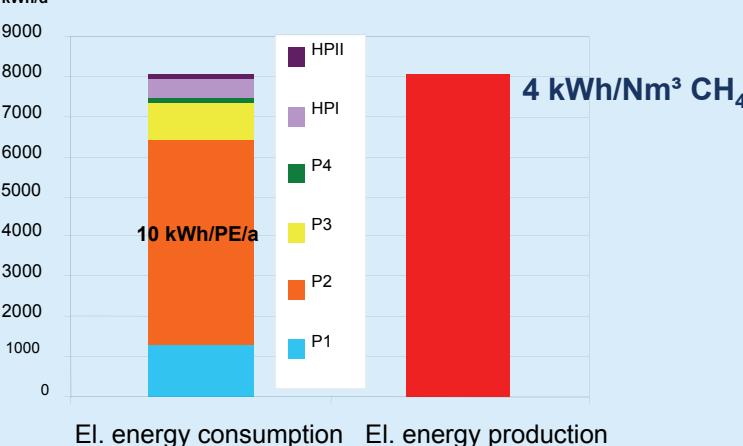
- 4 main processes without influent pumping
 - ⇒ pretreatment (screening, grit)
 - ⇒ mechanical-biological treatment
 - ⇒ sludge thickening, stabilisation
 - ⇒ Sludge dewatering, disposal
- 2 support processes (monitoring, other)
- All costs are related to 1 PE_{COD}
- Data quality assessment by mass balances for COD, P
- Process indicators for energy efficiency

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Benchmarking result 1st energy self-sufficient WWTP

170.000 PE

kWh/d

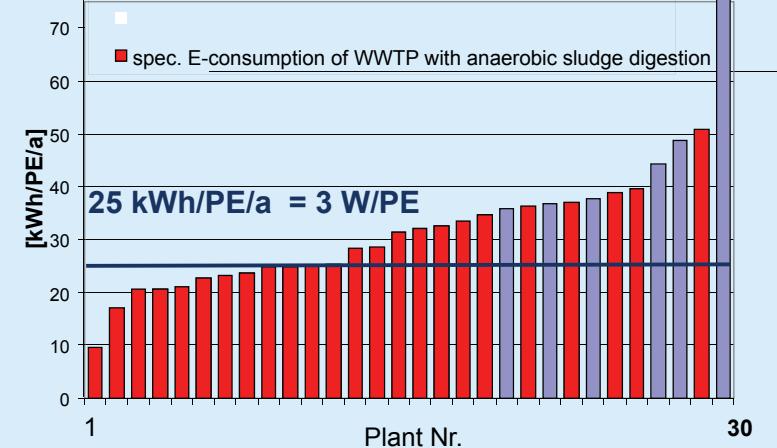


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Electric energy consumption (kWh/PE/a)

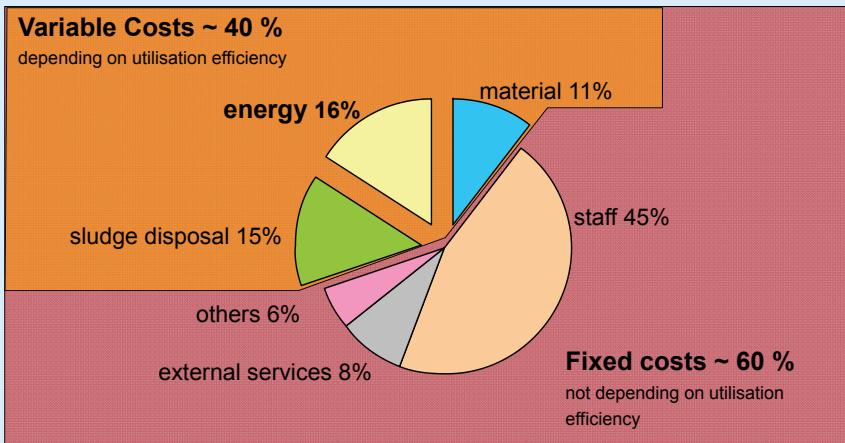
spec. E-consumption of WWTP with anaerobic sludge digestion

25 kWh/PE/a = 3 W/PE



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Distribution of operational costs



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Benchmarking results

- Energy demand of most of the large plants: 20 to 30 kWh/PE/year (~3W/PE)
- Aeration: 60 to 70 %
- Energy costs 15 to 17 % of operating costs
 - ⇒ Other cost factors: staff, sludge handling and disposal, monitoring
- Capital costs similar to operating costs
- Total costs 20 to 30 €/PE/a
- Actual energy costs 1.5 to 2 €/PE (1€ = 136 Yen)

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On which scientific background can we rely?

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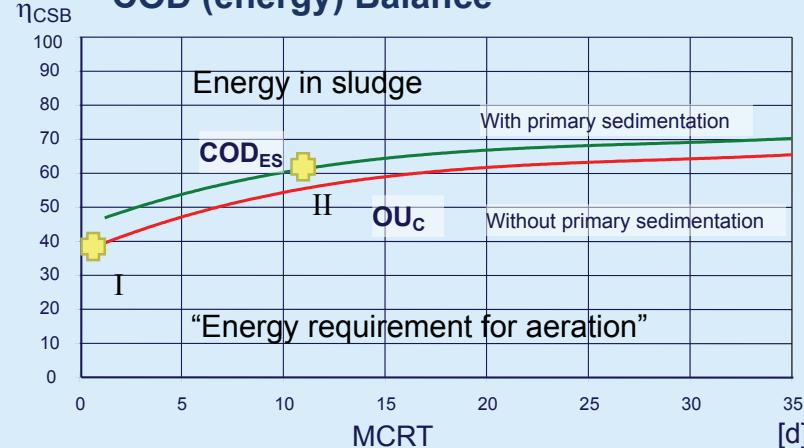
Important basics for this presentation

- COD allows mass and energy balances for organic waste water pollution and sludge treatment processes (1. law of TD)
- **Energy data are related to the yearly mean COD load of raw WW influent after grit removal expressed as**
- **PE = 120 g COD/d (60g BOD₅/d)**
- **1 kg COD corresponds to ~14 MJ (MWs)
 $120 \times 14 / 86.4 = 19 \text{ W/PE}$**

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Hard facts WWT COD (energy) Balance



Sum of OU_c und COD_{ES} = COD load removed

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Hard facts nitrification, N-removal

- 1 g N removed as N_2 = 1.7 g O_2 (DN, DA) OUDN
- 1 g NO_3-N in effluent = 4.6 g O_2 OUN
- $TN_{in} - TN_{effl} - TN_{sludge} = N_{removed}$

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Hard facts sludge digestion

- $COD(PS+ES) - COD(\text{digested sludge}) [\text{kg/d}] = COD \text{ of } CH_4$
- 1 kg COD = 350 Nm^3 Methane (CH_4) (Biogas)
- $TN(PS+ES) - TN(\text{dewatered sludge}) = TN(\text{reject water})$

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Vienna: EOS-Project

A new process scheme for waste water treatment with 8 years of full scale operational data and 1 year of on site pilot scale investigations

The Vienna Main Treatment Plant EOS-Project start of operation 2020)

Extension to energy producing operation,
start of full scale operation in 2020

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Vienna Main treatment Plant 1980
raw sludge incineration 3 Mio PE



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Vienna Main treatment Plant 2005
raw sludge incineration (4 Mio PE)



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Actual situation



Vienna Main Treatment Plant 2005

Designed for
4 Mio PE (240 t BOD₅/d)

Special 2-stage activated sludge process with
primary sedimentation for nutrient removal
raw sludge incineration with ash disposal

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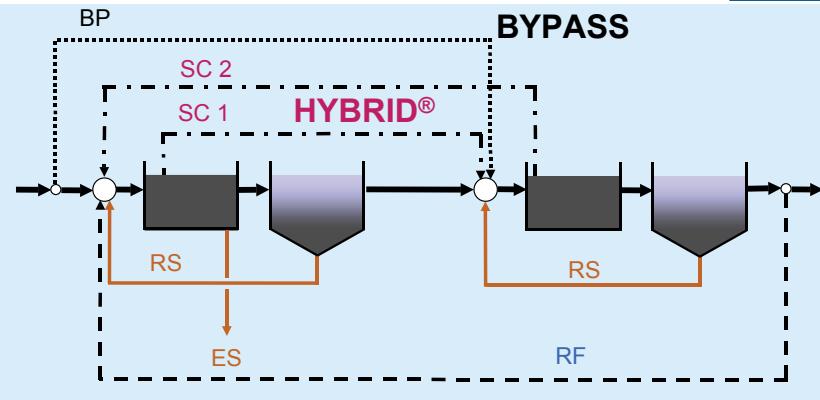
- Specific local situation:
 - ⇒ Temperature range from 10 to 23 °C (mean ~15°C)
 - ⇒ Combined sewer system
 - ⇒ Influent COD 650 mg/l
 - ⇒ Max. RW flow 18 m³/s (design flow)
 - ⇒ Max. dry weather flow actually 7 m³/s
 - ⇒ Mean yearly precipitation 550 mm

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- Treatment efficiency requirements: >75%N rem.
⇒ EU-UWWD Sensitive area, $\text{NH}_4\text{-N} < 5 \text{ mg/l}$, $\text{TP} < 1 \text{ mg/l}$

	2-stage	1-stage
PST-Volume:	5 l/PE	5 l/PE
Aeration tank volume I:	10 l/PE	130 l/PE
SST-Volume I:	18 l/PE	
Aeration tank volume II:	42 l/PE	
SST-Volume II	51 l/PE	51 l/PE
Total	126 l/PE	186 l/PE

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Assumptions for energy balance (based on full scale experience)

COD removal primary sedimentation:	30%
COD removal TP	92%
Aeration efficiency	2.0 kg O₂/kWh
COD of digested sludge	30 g/PE _{COD}
N-removal yearly mean	80%
N in sludge disposed	2 g N/PE
Gasmotors el. efficiency	4 kWh/Nm³ CH₄

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Case 1: 1-stage ASP with PS + sludge digestion

N/COD = **10/120**

COD Influent (N-influent 10 g N/EW/d, 80% rem.)	120 g/PE/d
COD Effluent primary sed. (COD-removal PS = 30%)	84 g/PE/d
COD in primary sludge	36 g/PE/d
COD Effluent TP	10 g/PE/d
COD removal in aeration tank: 84 – 10 =	74 g/PE/d
OUC (60% of COD removed)	44 g/PE/d
COD excess sludge (40% of COD removed)	30 g/PE/d
COD input digester: 36 + 30 =	66 g/PE/d
COD in digested sludge	30 g/PE/d
COD of digester gas production (CH ₄) 66 - 30 =	36 g/PE/d
OUN (denitrified N-load 6 g N/PE/d): (10-2-6.0)*4.6	9.2 g/PE/d
OUDN: 6.0*1,7 =	10,2 g/PE/d



Case 2: 2-stage ASP with PS sludge digestion N/COD = 10/120



COD influent (N influent 10 g N/PE/d, 80% rem.)	120 g/PE/d
COD effluent PS (COD removal by PS: 30%)	84 g/PE/d
COD of primary sludge	36 g/PE/d
COD effluent	10 g/PE/d
COD removal aeration tank 84 – 10 =	74 g/PE/d
OUC (40% 1.stage/ 60% 2.Stufe)	38g/PE/d
COD in excess sludge	36 g/PE/d
COD in digester feed: 36 + 36 =	72 g/PE/d
COD in digested sludge (N: 4 g/PE/d)	30 g/PE/d
COD in digester gas: 72 - 30 =	42 g/PE/d
OVN (nitrate in effluent 2g N/PE/d): (10-2-6) * 4,6	9.2 g/PE/d
OVDN: (4+2") * 1,7 =	10.2 g/PE/d

**Aerobic denitrification of reject water
(Anamox, Demon,....)**



Energy balance comparison

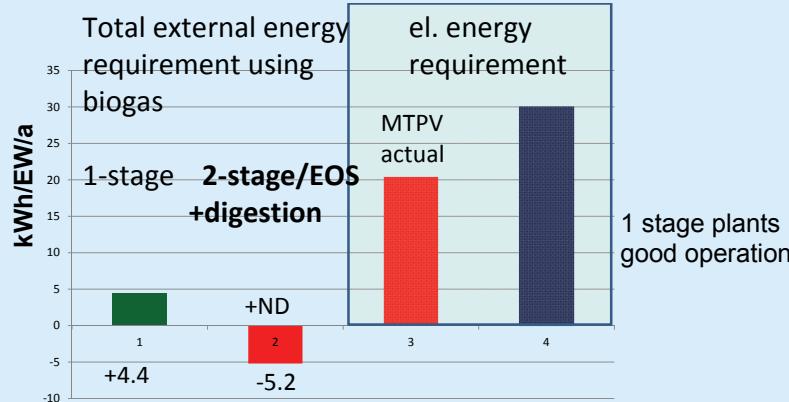


	Dim	1-stage $\eta_N = 80\%$	MTPV/EOS $\eta_N = 75\%$	HKA actually
"aeration efficiency"	kgO ₂ /kWh	2,0	2,0	Raw sludge
$\eta_{el\ gasmotor}$	%	38	38	Incineration
Power for aeration	W/EW	1,6	1,25	2,26 to 2,33
	W/PE	0,80	1,10	
Biogas el. power prod	W/EW	1,9	2,75	-
Total el. power requ.	W/EW	+ 0,5	- 0,4	2,3
El. Energy requ.	kWh/EW/a	+ 4,4	- 3,5	20,4

EOS Project (2020): MTPV with digestion, 83 % N-removal, reject water nitritation+Deni in AT 1



Energy requirements for nutrient removal plants in kWh/pe/a



Energy balance



	Aeration energy	Other energy	Total energy	$\eta_{el\ Gasmotor} 38\% \text{ el. efficiency}$
				W/PE
1-stage ASP N/COD = 8/120	1.6 +	0.7 =	2.3	1.9 W/PE
2-stage ASP N/COD = 8/120	1.2 +	0.9 =	2.1	2.7 W/PE

Power requirement production



EOS - project



- Reconstruction of 1. stage PS +AS
- Sludge Digestion with high solids concentration
- Reject water Nitritation-Denitritation
- Gas motors for electric energy production
- (Drying and Incineration of Digested sludge)
- (Recovery of Phosphorus from Ashes)

The plant remains connected to the public electric grid.
Independent operation would markedly increase the investment costs.

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Vienna main Treatment plant in 2020



Die Hauptkläranlage Wien im Jahr 2020. © LBS/OPEN Brand Design



Project EOS Pilot investigations

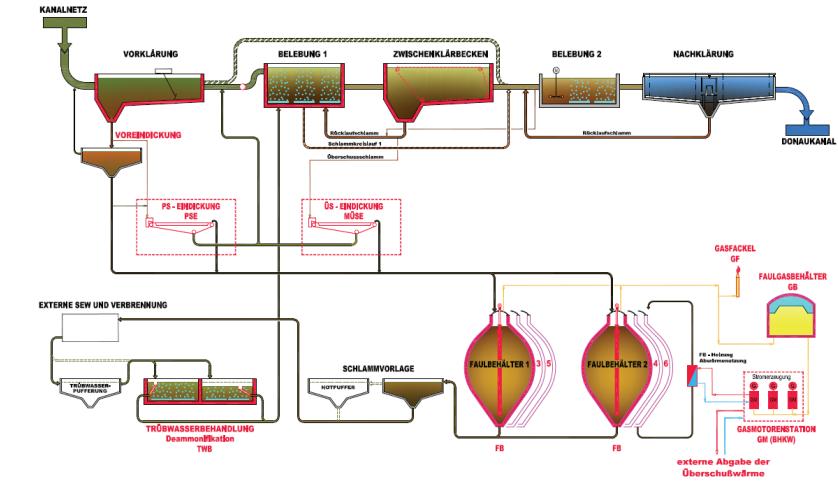
Markus Reichel, Helmut Kroiss
Institute for Water Quality , Ressource and Waste Management
Vienna University of Technology



Project EOS: Introduction



2020 process scheme for Main Treatment Plant of Vienna





- Questions to be answered
 - Digestion: design basis
 - Thickening behaviour, polymer requirements
 - COD and oDS-removal, gas production, gas composition
 - Operational experience
 - Rheology: viscosity, mixing behaviour, hydraulic losses
 - Reject water pretreatment: design basis
 - N-removal efficiency with real reject water
 - Process selection
 - Operational experience
- Pilot investigations
 - Pilot plant: Sludge digester, reject water treatment plant
 - Field investigations: Rheology, hydraulic losses
 - Lab scale experiments: ammonia inhibition of digestion, reject water treatment, rheology

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- Digestion:
- Mechanical sludge thickening
- Reactor volume 130 m³
- Gas motor: 15 kW_{electric}

- Full scale: 70.000 m³



- Reject water treatment (pretreatment)
 - Sludge dewatering with screw press
 - 5 m³ buffer tank
 - 2 reactors 2 m³ each



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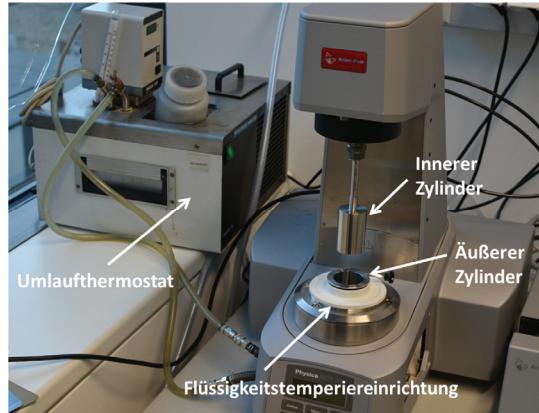


- Sludge digestion with high NH₃
- 3 Reactors 3 liters each
 - 35 to 40°C
 - Gas production: quality, quantity
 - Ammonia inhibition testing



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- Lab scale experiments:
- Rheological behaviour of:
 - Water
 - Thickened raw sludge
 - Digested sludge



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- Rheological pipe investigations at pilot plant



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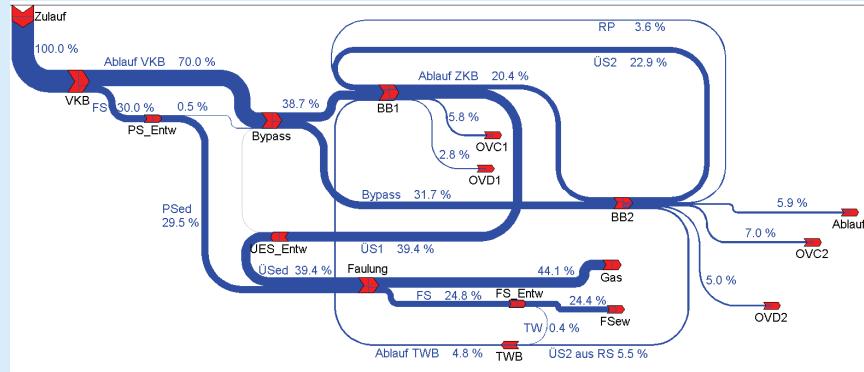
Parameter	Unit	Standard-operation	max. loading	high loading	Limits for digestion:	
					max. loading	High temperature
Ø temperature	°C	38,0	37,9	37,5	37,2	41,0
MCRT	d	24,2	19,8	17,3	11,5	23,0
pH	-	7,52	7,37	7,38	7,33	7,42
Ø DS _{Digester}	kg/m ³	40,65	38,63	43,02	44,73	43,71
Ø NH ₄ -N	mg/L	1.885	1.819	1.902	1.888	1.927
DS-removal	%	37,2	41,0	37,9	36,1	35,4
oDS removal	%	50,5	52,7	48,0	48,5	48,2
Ø oDS _{feed}	%	75,2	81,1	78,8	77,5	76,7
Ø oDS _{DS}	%	59,3	65,0	66,0	62,5	61,5
Ø CH ₄ content	%	64,6	63,2	67,3	68,3	67,3
COD-removal	%	59,4	59,3	58,5	52,0	75,1
COD/oDS _{in}	-	1,88	1,82	1,84	1,84	1,88
COD/oDS _{out}	-	1,60	1,60	1,44	1,65	1,60

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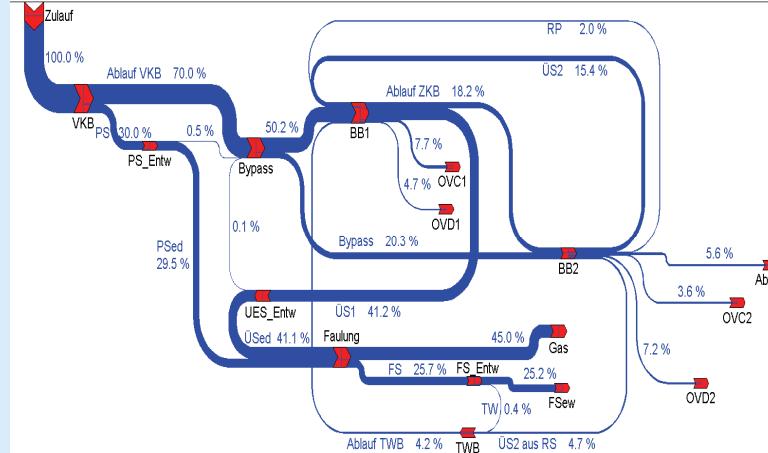
Simulation results for 2020

- **Data basis:** mean daily loading conditions for 365 days (COD, TN, TP influent and effluent, T, DS_{AT}, PS- and ES prod.) from the full scale plant (last 5 years data)
- **Modelling tool:** Site specific linked dynamic model for the 2-stage AS plant, sludge digestion, reject water treatment adapted to the data from full scale operation and pilot investigation data (basis: ASM 1, ADM)
- **Design load simulation:** Multiplication of the actual loading situation with a constant factor to simulate design loading (4 Mio. PE)
- Design Load: 85%ile of daily loadings (COD, TN) over one year

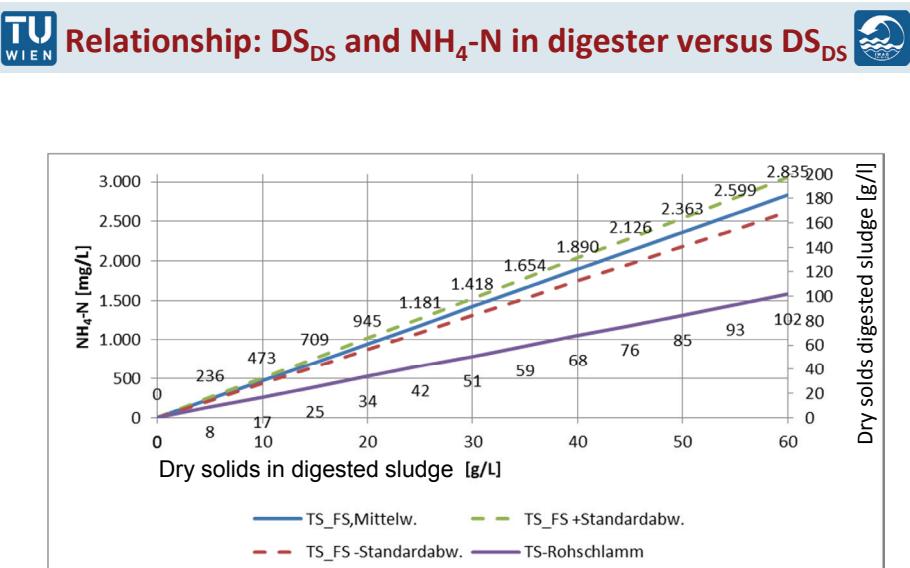
COD Balance with fixed bypass operation, nitritation/denitritation of reject water

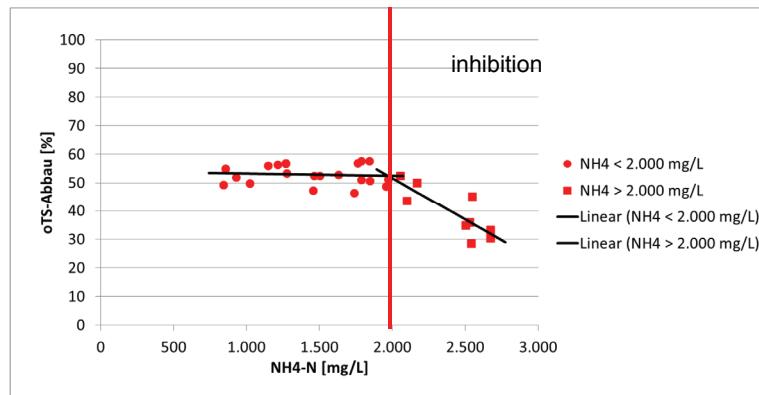


COD Balance for controlled bypass operation, nitritation/denitritation of reject water



Ammonia inhibition of anaerobic digestion

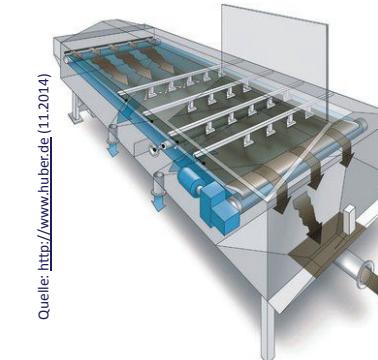
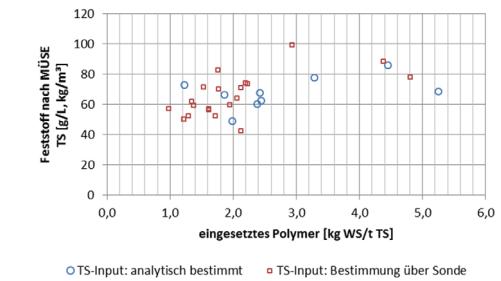




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Polymer addition versus DS_{thickened sludge}

- No good correlation possible

Quelle: <http://www.huber.de/11.2014>

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Digester mixing with gas production 2.7 Wh/Nm³/m depth

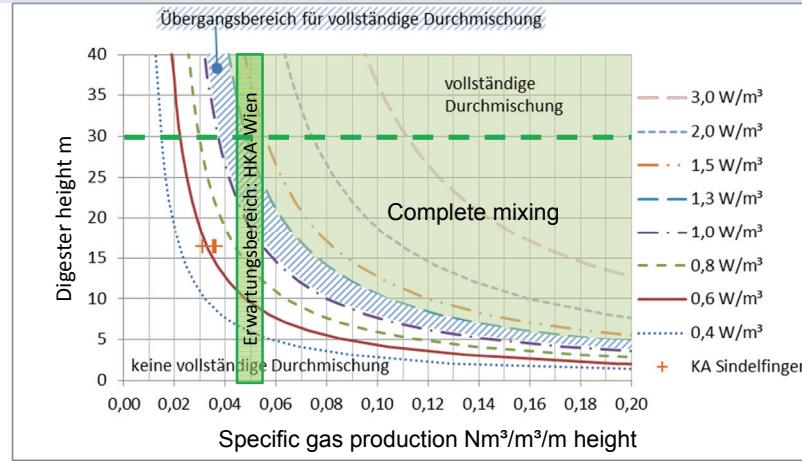
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Digestion needs complete mixing

- Power requirements for complete mixing:
 - Influencing factors: Mixing system, digester volume and shape, viscosity, inhibitory effects
 - Literature data: Mechanical mixers $2\text{--}6 \text{ W/m}^3$
Gas mixing $5\text{--}10 \text{ W/m}^3$
- Mixing by gas production
 - Advantage: no external energy requirement
 - Disadvantage: mechanical mixing equipment necessary for start up and operation
- Goal of investigations: will complete mixing need external energy supply under normal operating conditions

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Vienna MTP:

Gas production $0,043 \text{ m}^3_{\text{gas}} / (\text{m}^3_{\text{Digester}} \cdot \text{h})$...normal operation
 $0,055 \text{ m}^3_{\text{Gas}} / (\text{m}^3_{\text{FB}} \cdot \text{h})$design load

Digester height: 30 m

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- All sludges
 - Viscosity higher than water
 - Non-Newtonian behavior
- Calculation of hydraulic losses:
 - Mathematical models for non-newtonian liquids are existing but need parameters from experimental investigations
 - Influencing parameters: DS, Temp., type of sludge (PS/ES/DS, static/flowing, Polymer addition, etc.)
- Goal for design purposes:
 - Method for the determination of friction losses in pipes for different sludges and conditions:

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Rheological investigations

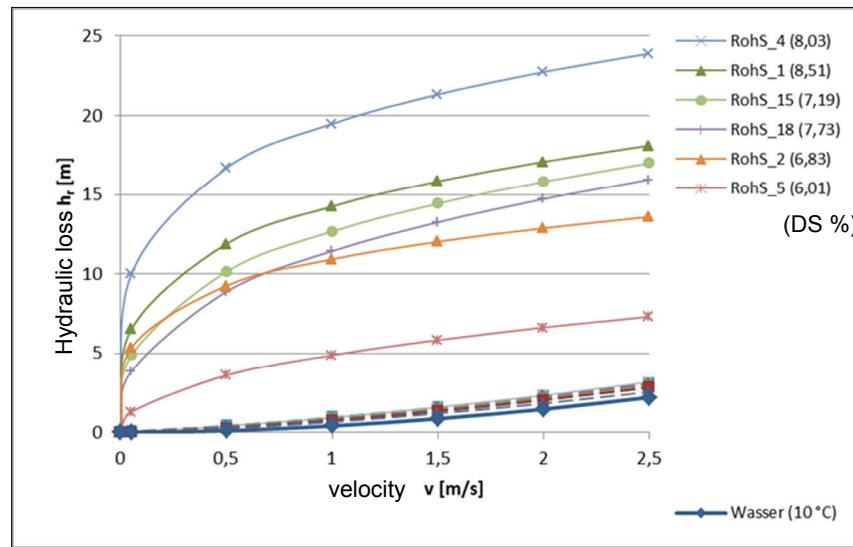
Findings:

- Hydraulic losses: thickened raw sludge>> digested sludge> water
- Raw sludge has strong non-newtonian behaviour
- Digested sludge similar to water
- DS concentration has dominating effect
- Quality of sludge (mixing of different sludges and seasonal effects) has great influence and cannot directly be transferred from one plant to another
- But estimation possible

Example:

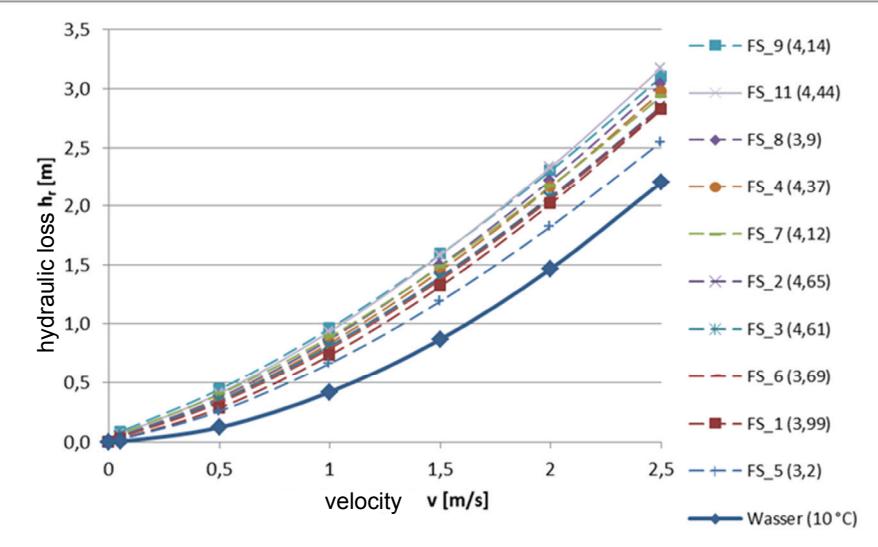
Pipe diameter 0,2 m, length 100m
 head loss versus velocity

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Marcus Reicher Thesis 2015

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Marcus Reicher Thesis 2015

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Treatment of reject water from sludge dewatering

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Nitrogen removal oxygen requirement

- Oxidation of 1 g of $\text{NH}_4\text{-N}$ to N_2 needs 1.7 g of O_2 irrespective of process used (deammonification or denitrification, denitration)
- If oxygen uptake for carbon removal is the limiting factor for denitrification, nitrate effluent will increase
- 1 g of $\text{NO}_3\text{-N}$ in the effluent needs additional energy for aeration of 2.9 g of oxygen

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Side stream nitritation /main stream denitritation in 1-stage ASP evaluation

- Ammonia concentration inhibitory for nitritation (NOB)
- Ammonia oxidisers (AOB) have higher growth rate at temperatures > 30°C than nitrite oxidisers (NOB)
- MCRT (2-5d) for stable operatio much less than for deammonification in side stream
- no solids removal before nitritation necessary
- Very low sensibility to operational conditions, fast start up
- As nitritation (AOB) reduces pH **alkalinity addition** (lime) will be necessary to achieve >50% N conversion to nitrite (goal ~80%)

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Deammonification evaluation

- Optimum ratio NO₂-N/NH₄-N = 1,32 difficult to achieve
- Anammox bacteria have a very low growth rate (sludge retention and solids removal from RW necessary,)
- Buffer tank for RW necessary
- Higher sensitivity and cost for control equipment
- If biocoenosis is affected slow recovery (low growth rate)
- Sludge transport for start up for a 2 mio PE plant not possible
- Process more sensitive as compared to nitritation denitritation (minimisation of nitrate production)
- If problems occur rapid reaction necessary (nitrate!)
- Higher personal costs for control and operation

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Nitritation in Pilot plant

- Test program
 - Adaptation of AS to nitritation by continuous increase of loading
 - Simulated steady state process (intermittend feeding)
 - Variation of sludge age (1,5 to 4 days)
 - Variation of temperature (32 to 45 °C)
 - Different control strategies for aeration (DO concentration and/or pH)
 - Only aerobic or intermittend aeration
 - Testing of probes (pH, T, NH₄-N, NO₃-N, O₂)
 - Foam development at RW treatment

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Nitritation in der Versuchsanlage

- Comparison SBR versus continuous flow
- Maximum conversion to nitrite by increasing pH by addition af alkalinity, economic evaluation
- RW-treatment at temperatures< 30 °C
- Minimum sludge age
- Shock load behaviour
- Influence of suspended solids (digested sludge) on nitritation process
- ...

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Nitritation results from pilot plant

- Nitritation is a very stable process (much less sensitive than deammonification)
- No nitrate production even at high sludge age
- ratio NO₂-N/NH₄-N without pH control at 1,10 to 1,15
- Slow changes of temperature do not affect the process
- Sudden temperature changes cause adaptation problems
- Stable operation at temperatures from 32 to 45 °C
- Stable operation at sludge ages from 1,5 to 4 d
- Foam development (h~15cm) normal
- Stable process under varying loading and temperature conditions as in reality

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Consequences for EOS project

- Decision is made to use nitritation in side stream and denitrification in 1. step ASP
- Both processes SBR and continuous flow will be possible (sludge age is equal to hydraulic detention (in SBR mode sludge age can be higher than hydraulic detention time))
- Chemical addition (lime) for pH control will be installed in order to control the nitritation efficiency (80% conversion of ammonia to nitrite probably the economic optimum)
- COD will not become limiting for denitrification in 1. step ASP even at 40% bypass ⇒ 85% N-removal from RW
- The nitritation volume can also be used for controlled discharge of nitrite to 1. step ASP in order to minimise aeration peaks for OU_C during the day



Conclusions

- Design and energy calculations as well as modelling can be based on sound theoretical background if enough full scale and pilot scale experience is available
- Process selection, efficient aeration and gas conversion are decisive for achieving energy “neutral” nutrient removal WWTP
- Life cycle analysis indicates that anaerobic digestion with gas-motors for energy recovery have nearly the same greenhouse gas emission (CO₂, N₂O, CH₄) as aerobic sludge stabilisation process. The differences mainly result from the necessary assumptions or selection of literature data (*Parravicini 2014*)

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Many thanks for your attention

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